ORIGINAL PAPER



Modeling gas-rock-water interactions in carbon dioxide storage capacity assessment: a case study of Jurassic sandstones in Poland

K. Labus · R. Tarkowski · M. Wdowin

Received: 20 February 2014/Revised: 20 June 2014/Accepted: 9 July 2014/Published online: 8 August 2014 © The Author(s) 2014. This article is published with open access at Springerlink.com

Abstract This paper presents the results of the modeling of gas-rock-water interactions for reservoir formation of one of potential geological carbon dioxide storage sites in the Polish Lowlands (Chabowo Anticline). The modeling performed with the use of a geochemical simulator was preceded by mineralogic-petrographical and petrophysical analysis of sandstones sampled together with pore water at the depths of 856-857 and 1,240-1,241 m. The study focused on determining the changes in rock matrix and reservoir parameters that take place as a result of carbon dioxide injection, and on the assessment of the formation suitability for carbon dioxide storage. The results have shown that in the analyzed gas-rock-water systems, in the modeled period of 20,000 years, the minerals able to trap carbon dioxide are calcite and dolomite. Mineral-trapping capacity of carbon dioxide, calculated for the sandstones considered, equals to 9.19 and 12.0 kg per one cubic meter of the formation.

 $\begin{tabular}{ll} \textbf{Keywords} & Carbon dioxide storage \cdot Hydrogeochemical \\ modeling \cdot Gas-rock-water interactions \cdot Sandstone \\ aquifer \\ \end{tabular}$

K. Labus (⊠)

Institute for Applied Geology, Silesian University of Technology, 2 Akademicka St., 44-100 Gliwice, Poland e-mail: klabus@o2.pl

R. Tarkowski · M. Wdowin Mineral and Energy Economy Research Institute of the Polish Academy of Sciences, 7 Wybickiego St., 31-261 Kraków, Poland

Introduction

Geological storage of carbon dioxide (CO₂) is one of the options of greenhouse gas mitigation as it reduces the release rate of CO2 to the atmosphere (e.g., Holloway 2004). After the injection of CO₂ into the subsurface, this gas can be trapped by several mechanisms: structural and stratigraphic trapping, residual CO2 trapping, solubility trapping and mineral trapping (Bachu et al. 2007). In mineral trapping, CO2 is bound by precipitating new carbonate minerals, and the storage potential depends on the temperature, pressure and the characteristics of aquifer formation, such as the mineralogy of the reservoir rock and the chemical composition of the brine. The kinetics of potential trapping reactions is also controlled by the specific surface of the minerals (the contact between skeletal grains and brine with dissolved CO₂) and the flow rate of fluids through the pore space. With regard to chemical reactions, the siliciclastic rocks (in particular calcium-, magnesium- or iron-rich reservoirs) are favored over carbonate ones, as they present higher capacity of CO₂ mineral trapping (Gunter et al. 1993). Mineral trapping is described as the safest mechanism in long-term storage of carbon dioxide (Metz et al. 2005); however, the time required for mineral precipitation is considerable, ranging from tens to thousands of years (Perkins and Gunter 1995). As a result of solubility trapping, CO₂ is dissolved in the formation water (Koide et al. 1992). This phenomenon is strongly time-dependent (Bachu et al. 1994). On the other hand, CO₂ solubility increases with pressure and decreases with temperature and salinity of the formation water. Dissolving CO₂ into the formation water triggers the process of convective mixing, which is much more efficient than diffusive mixing, and crucial for solubility trapping, as it



accelerates the overall dissolution of the CO₂ into the pore water.

Saline aquifers are promising for geological storage of CO₂ because of their storage potential and limited suitability for other uses (Hendriks and Blok 2004; Koide et al. 1992; Tarkowski and Wdowin 2011; Wdowin et al. 2013).

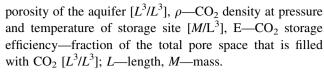
Geochemical monitoring of deep underground fluids is a basic procedure used to investigate the variations in fluid composition caused by CO₂ injection and water-rock interactions, and to track the fate of the injected CO₂ plume. Furthermore, the geochemical results are applied to carry out numerical modeling of deep geochemical processes (through thermodynamic database, saturation indexes, activities and speciation calculations of liquid phase) to evaluate theoretical reservoir equilibrium conditions among several phases and to reconstruct physico-chemical variations of different phases at non-equilibrium conditions on the basis of kinetically controlled reactions (e.g., Cantucci et al. 2009; Marini 2007). Therefore, in general, geochemical modeling is a helpful tool to assess the geochemical behavior of CO₂ in saline aquifer, including its dissolution in brine, and interactions with minerals, etc. (Trémosa et al. 2014), which allows to determine the suitability of geological structures for CO₂ storage (Labus et al. 2010). So far, numerous software packages were used for this purpose: e.g., PHREEQC (see: Tarkowski 2010; Van Pham et al. 2012), TOUGHREACT (Okuyama et al. 2013; Xu et al. 2006), CHILLER, SOLVEQ (Reed 1998; Rosenbauer et al. 2005), Geochemist's Workbench (Labus and Bujok 2011; Labus et al. 2010; Tarkowski et al. 2011; Zerai et al. 2006). The input data for geochemical modeling should include the following: formation temperature and pressure, petrophysical and petrological parameters of reservoir and cap rock, kinetic parameters of minerals, chemical data of pore fluids and the time of reaction.

In order to represent the different estimations made during calculations of carbon dioxide storage capacity, we can apply a techno-economic resource pyramid (Bradshaw et al. 2007) considering 4 technical and economic categories: total pore volume, prospective storage capacity, contingent capacity and operational storage capacity—characterized by different storage volumes and assessment certainty (CO2CRC 2008). Another version of the resource pyramid considers the relationships between trap types, trapping mechanisms and the time needed for a given trapping mechanism to be effective (Bradshaw et al. 2007).

A simplified method of calculation the CO_2 storage capacity in saline aquifers— G_{CO2} —was presented in DOE (2012):

$$G_{\text{CO}_2} = A \cdot h_{\text{g}} \cdot \phi_{\text{tot}} \cdot \rho \cdot E \tag{1}$$

where A—area of a region assessed for CO₂ storage [L^2], h_g —thickness of aquifer formation [L], ϕ_{tot} —average total



Bachu et al. (2007) proposed a much more precise calculation method, based on the assumption, that the trapping mechanisms are interdependent, and their effects should be combined to obtain the value of carbon dioxide trapping. The following components of storage capacity were distinguished according to this methodology: in structural and stratigraphic traps, in residual-gas traps, in solubility traps, in hydrodynamic traps and in mineral traps.

This paper presents the results of CO₂ storage capacity evaluation in solubility and mineral traps, based on geochemical modeling for Chabowo Anticline—one of important, potential geological formations suitable for underground storage of carbon dioxide in the Polish Lowlands. The Chabowo structure is of interest to Dolna Odra power plant (electric power installed 1,742 MWe), located 50 km away from it, and to Gorzów Heat and Power Plant (97.5 MWe and 220 MWt) at 100 km distance. The input data for the modeling were obtained by petrophysical and mineralogic-petrographical analysis as well as by pore water analysis of the reservoir rock of Chabowo Anticline.

The research was conducted in 2013, within the framework of collaboration between the Institute for Applied Geology, Silesian University of Technology—Gliwice, and the Mineral and Energy Economy Research Institute of the Polish Academy of Sciences—Kraków.

Materials and methods

Chabowo Anticline is located approximately 15 km ESE of Gryfino—Fig. 1. It was explored by three boreholes: Chabowo 1, Chabowo 2 and Chabowo 3, and by semidetailed reflection seismic survey. Geochemical modeling for the structure was performed based on 2 sandstone (reservoir rock) samples, taken at the depths of 856-857 and 1,240-1,241 m b.s.l. from Chabowo 1 and Chabowo 3 wells, respectively (Table 1). Analyzed samples of sandstone represent reservoir levels of the Lower Jurassic— Upper Pliensbachian (Komorowo Formation)—Fig. 2 (Tarkowski 2010). The percentage of sandstones in the Beds of Komorowo Formation ranges from 61 to 84 %. Their effective porosity reaches 21.55 % and permeability-540-800 mD. Pumping tests have shown discharge rates of 14.4–14.73 m³/h of brines, representing a chloride-sodium class I type, of TDS (total dissolved solids) within the range of 67-87.5 g/l, and the temperature ranging from 40 to 46 °C. Lower Jurassic Beds of Komorowo Formation are overlain by the Beds Gryfice





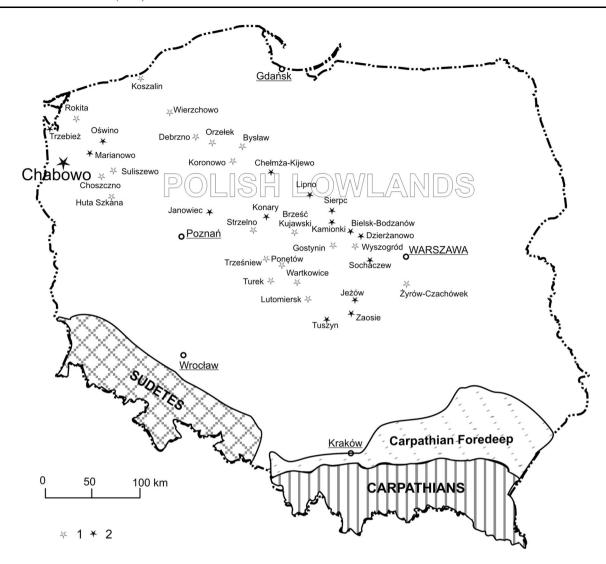


Fig. 1 Potential geological CO₂ storage sites in the Polish Lowlands. Explanations: *I* structures with one level of CO₂ storage, 2 structures with two levels of CO₂ storage

Formation (complex of claystones interbedded by sandy mudstone) (Tarkowski et al. 2014).

The analyzed reservoir rocks (Chabowo 1—856–857 m and Chabowo 3—1,240–1,241 m) are represented by moderately and poorly sorted, highly porous quartz sandstones. They are of medium and fine grain size 200–500 μm in diameter, but also bigger grains up to 1,500 μm can be found. Their contact-porous cement is composed mainly of clay minerals. Rock matrix typically contains quartz, lesser amounts of feldspars and micas. In sandstones from the deeper horizon, of 1,240–1241 m b.s.l, significant quantities of pyrite were also observed—Table 1.

Scanning electron microscopy and X-ray diffraction were used in the identification of mineral phases. The observations in transmitted light and planimetric analysis with JENALAB Karl Zeiss Jena polarizing microscope

were also made in order to assess the percentage composition of mineral assemblages.

X-ray diffraction analysis (XRD) was performed by means of a powder X-ray diffractometer Philips X'pert APD with PW 3020 goniometer and a copper lamp as well graphite monochromator. The analysis was carried out in the angular range of 5-65 20. For diffraction data processing, the Philips X'Pert software was used. The identification of mineral phases was carried out based on JCPDS–ICDD database.

To observe alterations in the rock matrix, the FEI Quanta 200 FEG scanning electron microscope, equipped with backscattered electron detector and EDS EDAX microprobe analyzer were used.

As the accuracy of image analysis could be insufficient in the case of porosimetric properties examination in sandstones (Labus 2001), the porosity and specific surface



Table 1 Basic parameters of rock samples from the Chabowo 1 and Chabowo 3 wells, and kinetic rate parameters at 25 °C (data from Palandri and Kharaka (2004)

	Sample no.	1		2	
	Well	Chabowo 1 856–857 1.5629 23.8		Chabowo 3 1,240–1,241 0.6968 17.0	
	Depth [m]				
	Mean capillary diameter [μm]				
	Porosity from porosimeter [%]				
	Kinetic rate k ₂₅ [mol/m ² s ⁻¹]	Composition [% vol.]	SSM [m ² /g]	Composition [% vol.]	SSM [m ² /g]
Quartz	1.02·e-14	68.2	4.53·e-6	62.8	7.55·e-6
Chalcedony	4.90·e-13	4.0	6.87·e-5	1.4	6.87·e-5
Illite	1.66·e-13	4.3	1.09·e-3	6.9	1.09·e-3
Annite	3.16·e-11	0.3	9.04·e-5	0.4	9.04·e-5
Muscovite	2.82·e-14	1.0	1.06·e-4	0.6	1.06·e-4
K-feldspar	3.89·e-13	8.0	4.69·e-6	5.8	7.82·e-6
Albite	2.75·e-13	4.0	4.59·e-6	2.9	7.45·e-6
Anorthite	7.59·e-10	4.0	4.35·e-6	2.9	7.25·e-6
Calcite	1.55·e-06	4.2	6.71·e-6	6.9	6.71·e-6
Goethite	1.15·e-08	2.0	4.26·e-6	1.4	4.26·e-6
Pyrite	2.82·e-05	_	_	8.0	7.98·e-6

SSM specific surface of analyzed minerals

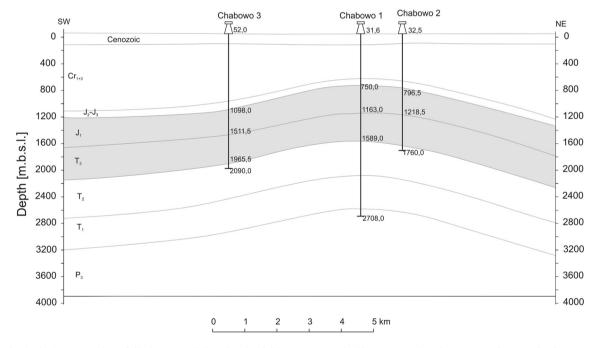


Fig. 2 Geological cross section of Chabowo Anticline (levels of CO₂ storage marked in gray. Explanations: *P* Permian, *T* Triassic, *J* Jurassic, *Cr* Cretaceous. (Tarkowski 2010)

were measured by means of mercury porosimetry. The AUTO PORE 9220 (Micromeritics) was applied at 30,000 psi maximum pressure, allowing for micropores detection.

Geochemical modeling, applying the methodology described in detail in Labus (2009) and Labus and Bujok (2011), was aimed at the assessment of storage capacity and the changes of rock matrix and reservoir parameters





Table 2 Groundwater chemistry

Parameter	Value
pH	7.0
Specific density [kg/l]	1.0586
TDS [g/l]	85.526
Cl ⁻ [mg/l]	50 700
SO_4^{2-} [mg/l]	728
HCO_3^- [mg/l]	110
Ca^{2+} [mg/l]	1 178
Mg^{2+} [mg/l]	164
Na ⁺ [mg/l]	29 500
K^+ [mg/l]	211

that could occur due to CO2 injection. The simulations were performed with the use of Geochemist's Workbench Pro 9.0. package in two stages: the first-simulation of immediate textural and mineralogical changes in the aquifer impacted by the beginning of CO2 injection

(100 days), the second—long-term effects of the storage (20,000 years). Kinetic rate constants, conforming to general Arrhenius-type rate equation for the minerals involved in modeled reactions (Table 1), were taken from Palandri and Kharaka (2004) and were recalculated for the formation temperatures: accepted as 40.0 °C for the depth interval 856-857 m, and as 46.0 °C for 1,240-1,241 m. Pore water composition used in the modeling is presented in the Table 2. It was also assumed that under the injection pressure (85 % of lithostatic pressure), CO₂ pressure reaches 176.8 and 258.0 bar, respectively. Gas fugacities f_{CO2} equal to 68.4 and 86.9 bar, were calculated after Duan et al. (2006), using online calculator of The Duan Group.

Specific surface area of minerals (SSM) required for the simulation of kinetic reactions was calculated based on spherical grains model, using the radius (calculated based on image analysis discussed in Labus 2001), molar volume and molecular weight of each of mineral, according to the following formula:

Fig. 3 Changes in: pH, quantity of carbonate minerals and porosity on the stage of CO2 injection—sample 1

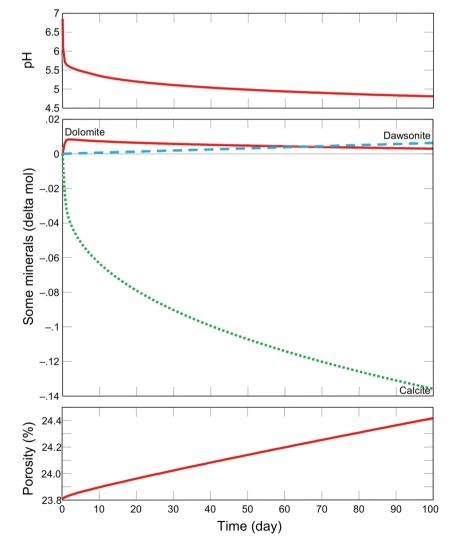
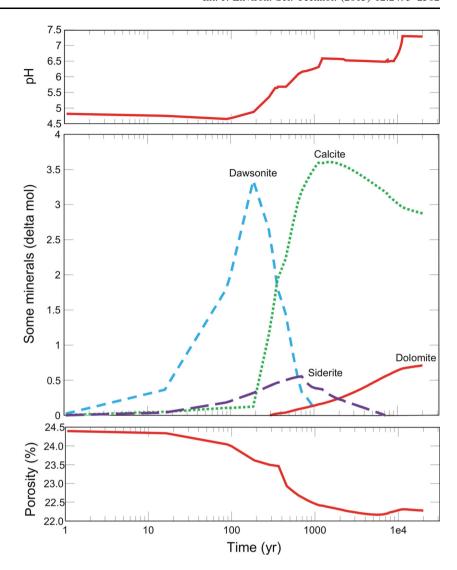




Fig. 4 Changes in: pH, quantity of carbonate minerals and porosity since the termination of CO₂ injection—sample 1



$$SSM = \frac{A \cdot v}{V \cdot MW},\tag{2}$$

where A-sphere area [m²], v—molar volume [m³/mol], V—sphere volume [m³] and MW—molecular weight [g/mol] of a given mineral phase. The values of the specific surface areas used in calculations are presented in the Table 1.

Results and discussion

The results of modeling for Chabowo 1 Well

During the first stage, lasting for 100 days, the CO_2 injection causes an increase of gas fugacity to the assumed value: f_{CO2} —68,4 bar. In effect, we can observe a significant elevation of CO_2 (aq) and HCO_3 ⁻ concentrations, to 1.296 and 0.076 mol/l, respectively, and a drop of pore water's pH to 4.82 pH (Fig. 3). The increase of porosity by

0.6 percent points (pp) to 24.4 %, barely influencing the injected fluid penetration into the aquifer, is mainly controlled by the dissolution of calcite. The changes in the quantity of carbonates at the stage of CO_2 injection are shown in the Fig. 3.

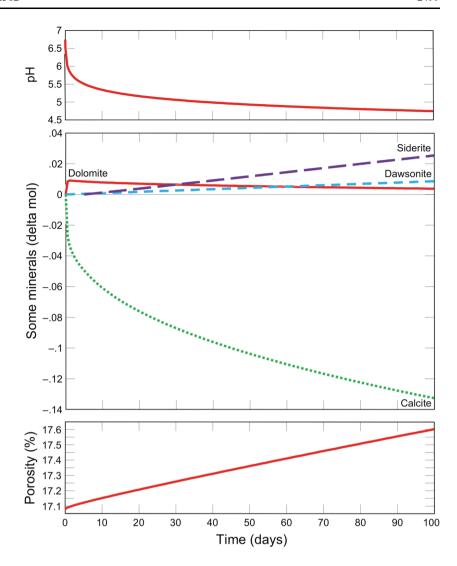
In the second stage, within a period of approximately 200 years, CO_2 fugacity drops to the value of about 1 bar. The $CO_2(aq)$ and HCO_3 concentrations fall significantly to $5.4 \cdot 10^{-6}$ and $9.4 \cdot 10^{-5}$ mol/l, respectively; rising of pH reaches the value of 7.29. The porosity falls to 22.3 % (1.5 pp lower compared to the primary porosity) causing a slight permeability damage of the aquifer (Fig. 4).

For the first 200 years of the second stage, the processes of dawsonite and siderite precipitation (Fig. 4) are mainly responsible for the porosity decrease. Subsequently, both these minerals undergo dissolution, triggering crystallization of calcite and dolomite, respectively. Finally, 2.87 mol of secondary calcite and 0.71 mol dolomite precipitate per unit volume of modeled aquifer rock—UVR—equal to





Fig. 5 Changes in: pH, quantity of carbonate minerals and porosity on the stage of CO₂ injection—sample 2



 0.01 m^3 , within 20 ka (20,000 years) of simulated CO_2 aquifer storage.

The results of modeling for Chabowo 3 Well

The gas—rock—water interactions were modeled for sample no. 2 as well. The main mechanisms playing the role during the first stage—the CO₂ injection—are similar to the ones working in the case of sample no. 1. The pH of pore waters drops to 4.75 pH (Fig. 5). We can observe siderite precipitation, which is connected with the dissolution of goethite and pyrite. A small increase of porosity is also noticed, mainly due to the significant volume of degraded calcite (Fig. 5).

The second stage is characterized by a sudden drop of fugacity, a stepwise decrease of $CO_2(aq)$ and HCO_3 concentrations, and a quick growth of pH (Fig. 6). The porosity decreases by 1.2 pp (to 15.8 %, compared to the primary value—17 %), in favor of the rock insulating properties. Calcite (1.79 mol per UVR) and dolomite

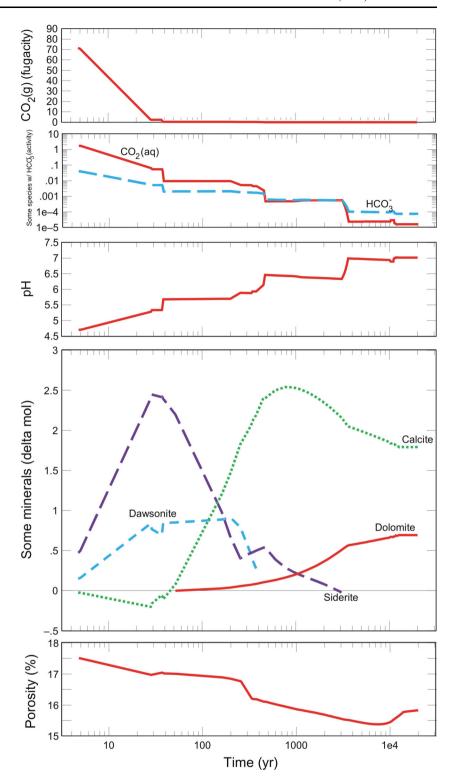
(0.69 mol per UVR) are the precipitating mineral phases that are able to efficiently trap injected CO₂ (Fig. 6). It should be noted that some amounts of siderite that were already formed at the stage of injection—about 0.2 mol, are completely degraded in the stage II.

Storage capacity calculations

The trapping capacity of the formations analyzed (Table 3) was calculated under the following assumptions: The unit volume of modeled aquifer rock—UVR—is equal to $0.01~\mathrm{m}^3$, and the primary porosity value (prior to the storage) is equal to n_p . Then, the rock matrix volume measured in UVR in $1~\mathrm{m}^3$ of formation amounts to $100(1-n_\mathrm{p})$. According to the modeled reactions, a certain quantity of carbonate minerals dissolves or precipitates per each UVR. On this basis, the CO_2 balance and eventually the quantity of CO_2 trapped in mineral phases is calculated. Modeled chemical constitution of pore water allows for the



Fig. 6 Changes in: f_{CO2}, concentrations of CO₂(aq) and HCO₃⁻, pH, quantity of carbonate minerals and porosity since the termination of CO₂ injection—sample 2



calculation of the quantity of CO_2 trapped in the form of the solution. After simulated 20 ka of storage, the final porosity is n_f . Pore space is assumed to be filled with pore water of known (modeled) concentrations of CO_2 -containing aqueous species: e.g., HCO_3^- , $CO_2(aq)$, CO_3^{2-} , $NaHCO_3$, etc. (expressed in molHCO $_3^-$ /I).

An example for No. 1 sample is given below: The primary porosity— n_p is about 23.8 %, thus 1 m³ of formation contains 76.2 UVRs. For each UVR, 2.87 mol calcite and 0.71 mol dolomite precipitate, trapping 3.58 mol of CO₂. After 20 ka of storage, the final porosity— n_f is equal to 22.3 %; therefore, 1 m³ of formation is assumed to contain





Table 3 Mineral and solubility trapping capacity of analyzed formations

	Sample no.	1	2
Porosity (%)	n _p –primary—0 ka	23.8	17.0
	n_f —final—20 ka	22.8	15.8
Precipitating	Calcite	2.87	1.79
Minerals [mol/UVR]	Dolomite	0.71	0.69
CO_2	mol/UVR	3.58	2.48
Mineral trapping*	$[kg CO_2/m^3]$	12.16	9.19
CO_2	mol/l	$1.8 \cdot 10^{-4}$	$4.2 \cdot 10^{-4}$
Solubility trapping*	$[kg CO_2/m^3]$	$1.77 \cdot 10^{-3}$	$2.92 \cdot 10^{-3}$
SUM [kg CO ₂ /m ³]		12.162	9.193

^{*} Explanations in the text

 $0.223~{\rm m}^3$ of pore water. The fluid contains $1.8\cdot 10^{-4}~{\rm mol}$ HCO₃⁻/l, therefore, the calculated solution trapping for the formation is $1.77\cdot 10^{-6}~{\rm kg}~{\rm CO}_2$ /l.

Mineral trapping is strongly dependent on the thermodynamic regime of the reservoir as well as on chemical composition of formation waters, flow rate of fluids, and mineralogical composition and reactive surface of the rock matrix (Gunter et al. 2004). The reactive surface size of a rock depends on grain and pore size distribution, and on the intensity of grain contacts. The flow rate may vary depending on hydraulic gradients, permeability of rocks and water viscosity (controlled by temperature and water mineralization, and the pressure to some extent). In the context of such complex processes involved and the precision of the data required to assess the amounts of secondary minerals precipitation, any exact regional-scale estimations of CO₂ storage capacity are impossible. Only small-area scale estimates of the amount of mineral-trapped CO₂, based on numerical simulations, supported by experiments and field data from natural analogs, are feasible. The CO₂ mineral-trapping capacity of unitary rock volume, within the modeled 20 ka period, is much more efficient than storage through solubility trapping; however, the mineral trapping process is significantly slower (e.g., Xu et al. 2003).

Conclusion

In the analyzed gas—rock—water systems, within the modeled period of 20,000 years, the minerals able to trap CO_2 are calcite and dolomite. Mineral-trapping capacity of 9.19 and 12.16 kg CO_2/m^3 (Table 3) calculated for the Chabowo Anticline sandstones is almost two to four times higher than the values described in Labus and Bujok (2011) for Carboniferous sandstone aquifers of the Upper Silesian Coal Basin. It is, however, comparable to the data presented by

Balashov et al. (2013), for sandstone, where 30 % out of 34.5 kg CO₂ stored per m³ of rock precipitated as calcite. Similar capacities, reaching 10.3 kg CO₂/m³, were also assessed for Jurassic aquifers of potential storage site within the Belchatów area (Labus et al. 2010), while for a glauconitic sandstone from Alberta Sedimentary Basin—Xu et al. (2004) reported 17 kg per m³ of formation.

Under the conditions of Chabowo Anticline sandstones, the solubility trapping performance (up to $2.92 \cdot 10^{-3}$ kg CO_2/m^3) is negligible compared to those of mineral trapping. Also the changes in rock porosity, observed due to the simulation, are insignificant in the context of permeability damage of the sandstones analyzed.

The presented results proved the occurrence of high storage capacity horizons within the analyzed area. More exact assessment of the geological space suitability for CO₂ storage requires, however, the determination of variability of petrophysical and petrological properties and accurate tectonic-structural analysis of the formation.

Acknowledgments This work was funded by the National Science Centre, according to the decisions No. DEC-2011/03/B/ST10/04340 and DEC-2012/05/B/ST10/00416. The authors are grateful to the three anonymous reviewers whose comments and suggestions helped improve and clarify this manuscript.

Open Access This article is distributed under the terms of the Creative Commons Attribution License which permits any use, distribution, and reproduction in any medium, provided the original author(s) and the source are credited.

References

Bachu S, Gunter WD, Perkins EH (1994) Aquifer disposal of CO₂: hydrodynamic and mineral trapping. Energy Convers Manag 35(4):269–279

Bachu S, Bonijoly D, Bradshaw J, Burruss R, Holloway S, Christensen NP, Mathiassen OM (2007) CO₂ storage capacity estimation: methodology and gaps. Int J Greenh Gas Control 1:430–443

Balashov VN, Guthrie GD, Hakala JA, Rimstidt CL, Lopano JD, Brantley SL (2013) Predictive modeling of CO₂ sequestration in deep saline sandstone reservoirs: impacts of geochemical kinetics. Appl Geochem 30:41–56

Bradshaw J, Bachu S, Bonijoly D, Burruss R, Holloway S, Christensen NP, Mathiassen OM (2007) CO₂ storage capacity estimation: issues and development of standards. Int J Greenh Gas Control 1(1):62–68

Cantucci B, Montegrossi G, Vaselli O, Tassi F, Quattrocchi F, Perkins EH (2009) Geochemical modeling of CO₂ storage in deep reservoirs: the Weyburn Project (Canada) case study. Chem Geol 265(1-2):181-197

CO2CRC (2008) Storage capacity estimation, site selection and characterisation for CO₂ Storage Projects. Cooperative Research Centre for Greenhouse Gas Technologies, Canberra. CO2CRC Report No. RPT08-1001, p. 52

DOE (2012) The North American Carbon Storage Atlas: Appendix B-summary of methodologies used to estimate CO_2 storage



- resource. http://www.netl.doe.gov/File%20Library/Research/Carbon%20Seq/Reference%20Shelf/NACSA2012.pdf. Accessed 12 June 2014
- Duan ZH, Sun R, Zhu C, Chou IM (2006) An improved model for the calculation of CO₂ solubility in aqueous solutions containing Na⁺, K⁺, Ca²⁺, Mg²⁺, Cl⁻, and SO₄²⁻. Mar Chem 98:131–139
- Gunter WD, Perkins EH, McCann TJ (1993) Aquifer disposal of CO₂-rich gases: reaction design for added capacity. Energy Convers Manag 34(9–11):941–948
- Gunter WD, Bachu S, Benson SM (2004) The role of hydrogeological and geochemical trapping in sedimentary basins for secure geological storage for carbon dioxide. In: Baines SJ, Worden RH (eds) Geological Storage of Carbon Dioxide. Geological Society Special Publication 233, London, pp 129–145
- Hendriks CA, Blok K (2004) Underground storage of carbon dioxide. Energy Convers Manag 34(9–11):949–957
- Holloway S (2004) Underground sequestration of carbon dioxide–a viable greenhouse gas mitigation option. Energy 30:2318–2333
- Koide H, Tazaki Y, Noguchi Y, Nakayama S, Iijima M, Ito K, Shindo Y (1992) Subterranean containment and long-term storage of carbon dioxide in unused aquifers and in depleted natural gas reservoirs. Energy Convers Manag 33(5–8):619–626
- Labus M (2001) Comparison of computer image analysis with mercury porosimetry in sandstone porosity measurement. Geol Q 45(1):75–79
- Labus K (2009) Modeling hydrochemical effects of carbon dioxide sequestration in saline aquifers of the Upper Silesian Coal Basin. Monograph of the Silesian University of Technology 235, Gliwice, p 106
- Labus K, Bujok P (2011) CO₂ mineral sequestration mechanisms and capacity of saline aquifers of the Upper Silesian Coal Basin (Central Europe)—modeling and experimental verification. Energy 36:4974–4982
- Labus K, Tarkowski R, Wdowin M (2010) Assessment of CO₂ sequestration capacity based on hydrogeochemical model of water-rock-gas interactions in the potential storage site within the Belchatów area (Poland). Miner Resour Manag 26(2):69-84
- Marini L (2007) Geological sequestration of carbon dioxide: thermodynamics, kinetics and reaction path modeling. Elsevier, Amsterdam
- Metz B, Davidson O, Coninck H, Loos M, Meyer L (eds) (2005) Carbon dioxide capture and storage. The IPCC Special Report. Cambridge University Press, Cambridge, p 442
- Okuyama Y, Norifumi TN, Munetake SM, Ajima S, Akasaka C (2013) Reactive transport simulation study of geochemical CO₂ trapping on the Tokyo Bay model—with focus on the behavior of dawsonite. Appl Geochem 30:57–66
- Palandri JL, Kharaka YK (2004) A compilation of rate parameters of water-mineral interaction kinetics for application to geochemical modeling. US Geological Survey. Open File Report 2004-1068, pp 1-64
- Perkins EH, Gunter WD (1995) Aquifer disposal of CO₂-rich greenhouse gases: modelling of water–rock reaction paths in a siliciclastic aquifer. In: Kharaka YK, Chudaev OV (eds)

- Proceedings of the 8th International Symposium on water-rock interaction, Balkema, Rotterdam, pp 895–898
- Reed MH (1998) Calculation of simultaneous chemical equilibria in aqueous-mineral–gas systems and its application to modeling hydrothermal processes. In: Richards J, Larson P (eds) Techniques in hydrothermal ore deposit geology. Reviews in economic geology 10:109–124
- Rosenbauer RJ, Koksalan T, Palandri JL (2005) Experimental investigation of CO₂-brine-reck interactions at elevated temperature and pressure: implications for CO₂ sequestration in deep-saline aquifer. Fuel Process Technol 86:1581–1597
- Tarkowski R (Ed.) (2010) Potential geological structures to CO₂ storage in the Mesozoic Polish Lowlands (characteristics and ranking). Studies, Theses, Monographs of the MEERI PAS 164, Kraków, pp 138
- Tarkowski R, Wdowin M (2011) Petrophysical and mineralogical research on the influence of CO₂ injection on mesozoic reservoir and Caprocks from the Polish Lowlands. Oil Gas Sci Technol Rev IFP Energy Nouv 66(1):137–150
- Tarkowski R, Wdowin M, Labus K (2011) Results of mineralogic-petrographical studies and numerical modeling of water-rock-CO₂ system of the potential storage site within the Belchatow area (Poland). Energy Proc 4:3450–3456
- Tarkowski R, Dziewińska L, Marek S (2014) The characteristics of selected potential geological structures for CO₂ underground storage in Mesozoic deposits of the Szczecin–Mogilno–Uniejów Trough. Studies, theses, monographs of the MEERI PAS, Kraków, pp 92
- Trémosa J, Castillo C, Vong CQ, Kervévan C, Lassin A, Audigane P (2014) Long-term assessment of geochemical reactivity of CO₂ storage in highly saline aquifers: application to Ketzin, In Salah and Snøhvit storage sites. Int J Greenh Gas Control 20:2–26
- Van Pham TH, Aagaard P, Hellevang H (2012) On the potential for CO₂ mineral storage in continental flood basalts—PHREEQC batch- and 1D diffusion-reaction simulations. Geochem Trans 13(5):2–12
- Wdowin M, Tarkowski R, Franus W (2013) Supplementary studies of textural and mineralogical changes in reservoir and cap rocks from selected potential sites suitable for underground CO₂ storage. Arab J Sci Eng. doi:10.1007/s13369-013-0862-0
- Xu T, Apps JA, Pruess K (2003) Reactive geochemical transport simulation to study mineral trapping for CO₂ disposal in deep arenaceous formations. J Geophys Res 108:2071–2084
- Xu T, Apps JA, Pruess K (2004) Numerical simulation of $\rm CO_2$ disposal by mineral trapping in deep aquifers. Appl Geochem 19:917–936
- Xu T, Sonnenthal EL, Spycher N, Pruess K (2006) TOURGHRE-ACT: a simulation program for non-isothermal multiphase reactive geochemical transport in variably saturated geologic media. Comput Geosci 32:145–165
- Zerai B, Saylor BZ, Matisoff G (2006) Computer simulation of CO₂ trapped through mineral precipitation in the Rose Run Sandstone, Ohio. Appl Geochem 21:223–240



